

## On the occurrence of the Asiatic cyprinid *Pseudorasbora parva* in the Netherlands

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A large scale inventory along the Meuse River (the Netherlands), showed that floodplain lakes function as spawning, nursery and adult habitats, while the main river channel merely serves as a dispersal corridor for adult *Pseudorasbora parva*, one of the most successful invasive fish species that have colonized Europe.

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The topmouth gudgeon *Pseudorasbora parva* (Temminck & Schlegel 1846) is native to East Asia, *e.g.* northern China, Korea, Japan and Taiwan (Bănărescu, 1999). It was first recorded in Europe in 1960, after its inadvertent introduction into fishponds along the River Danube in Romania (Bănărescu, 1964). From the Lower Danube region, this species has exhibited a rapid spread in all directions. It has dispersed north-west through the Danube and Rhine river systems, successively invading Hungary in 1970 (Biro, 1972), Czechoslovakia in 1974 (Zitman & Holčík, 1976), Austria in 1982 (Weber, 1984), East Germany in 1984 (Arnold, 1985), West Germany in 1987 (Lelek & Köhler, 1989), Poland in the early 1990s (Kotusz & Witkowski, 1998), the Netherlands in 1992 (Lenders, 1993), Flanders (Belgium) in 1992 (Vandelannoote *et al.*, 1998) and Denmark in 2002 (Olesen *et al.*, 2003). Around the late 1970s it was intentionally introduced into Lake Skadar (Albania), as a food source for predatory fishes reared in hatcheries (Wildekamp *et al.*, 1997). From here, *P. parva* subsequently spread to Serbian, Montenegrin, Macedonian and Greek waters, due to unintentional introductions (Bianco, 1988; Cakic *et al.*, 2004). *Pseudorasbora parva* has

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also been recorded in Italy (c. 1988; Sala & Spampanato, 1991), France (1980s; Allardi & De Chancerel, 1988), England (1990s; Gozlan *et al.*, 2002) and Spain (in 2001; Caiola & De Sostoa, 2002); its presence here has been attributed to human-mediated introductions, because natural spread to these countries is impeded by the presence of geographic barriers (*e.g.* the Alps, the Pyrenees and the North Sea). The species' presence has further been reported from countries south-east of the Lower Danube region, *e.g.* the Republic of Moldova (Blanc *et al.*, 1971), Ukraine (Movchan & Smirnov, 1981), the Russian Federation (Bogutskaya & Naseka, 1996), Azerbaijan (Reshetnikov *et al.*, 1997), Kazakhstan (Blanc *et al.*, 1971), Uzbekistan (Kamilov & Urchinov, 1995), Turkmenistan (Coad, 1995), Turkey (Wildekamp *et al.*, 1997), Iran (Coad & Abdoli, 1993), Afghanistan (Coad, 1981) and Algeria (Perdices & Doadrio, 1992).

*Pseudorasbora parva* is one of the most successful invasive fish species that have recently colonized Europe, reaching an almost pan-Eurasian distribution within a period of <40 years (Gozlan *et al.*, 2002). Its success can be explained by its opportunistic life-history characteristics (*e.g.* omnivore, short generation time, polyphilic spawner, high reproductive effort and parental care; Rosecchi *et al.*, 2001) and by its wide physiological tolerance, which enables it to withstand extreme environmental conditions (*e.g.* low water levels, elevated temperatures, low oxygen concentrations and algal blooms; Bănărescu, 1999; Pollux & Pollux, 2004). Although little information is available on the interaction of *P. parva* with native European fish species, negative impacts have been reported, including predation on the eggs and larvae of native species (Stein & Herl, 1986), interspecific competition for food (Declerck *et al.*, 2002) and introduction of exotic parasites (Gozlan *et al.*, 2005). Knowledge of the species' dispersal capability, habitat preference and the existing scale of invasion is therefore essential for predicting where, and which, fish communities may be potentially threatened by future invasions of *P. parva*.

*Pseudorasbora parva* was first discovered in the Netherlands in 1992, at the confluence of the Aalsbeek River with the Meuse River, near the city of Tegelen [one specimen, 25 mm total length,  $L_T$ , 11 April 1992; Lenders, 1993; Fig. 1(a)]. A few months later *P. parva* was caught in the Swalm River (Meuse catchment), near the city of Swalmen (five specimens, 20–40 mm  $L_T$ , 17 October 1992; Lenders, 1993). Although the exact method and route of invasion is unknown, *P. parva* is thought to have reached the Netherlands from Germany, *via* the Rhine drainage (Lenders, 1993). Although *P. parva* has gradually spread to other locations in the Netherlands since 1992, its centre of distribution still appears to lie in the south-eastern part of the country, along the Meuse River, in the Province of Limburg [De Nie, 1997; Lenders, 2000; Anon., 2005; Fig. 1(a)]. Unfortunately, most reported observations concern accidental catches of single specimens, predominantly at the confluences of the Meuse River and its tributaries. A systematic inventory assessing the full scale of the invasion, in terms of total numbers of specimens and geographic distribution, is, to date, largely lacking.

To assess population densities and potential habitat preference in Dutch waters, 35 locations were selected in the Meuse River (at the confluences of the Meuse River and its tributaries) and 10 floodplain lakes along the Meuse River [Fig. 1(b)]. These sites were sampled between May 2004 and August

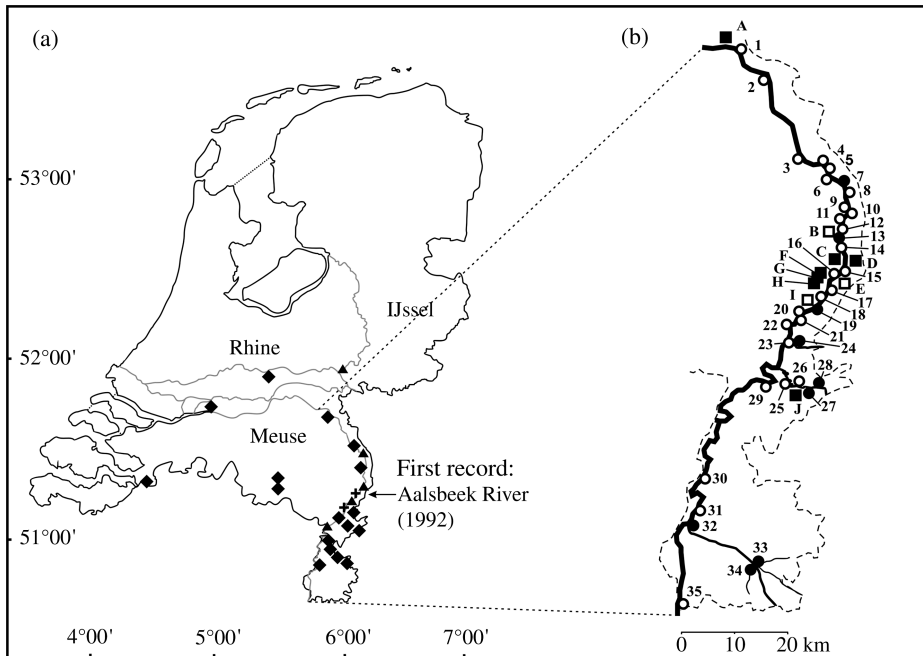


FIG. 1. (a) Geographic locations of recorded catches of *Pseudorasbora parva* in the Netherlands (+, 1992; ▲, 1993–1997; ◆, 1998–2005, catches from this study not included). (b) Study area showing the sampled river locations 1–35 (○) and floodplain lakes A–J (□). River locations (●) and floodplain lakes (■) where *P. parva* were caught are indicated.

2005 (river locations:  $n = 152$  times and floodplain lakes:  $n = 24$  times), by means of hand-nets (mouth size  $70 \times 50$  cm; mesh-size  $3.0 \times 3.0$  mm) (Pollux *et al.*, 2006). The results showed a highly significant difference in mean *P. parva* densities between river locations (mean  $\pm$  s.d.:  $0.11 \pm 0.61$  individuals  $100 \text{ m}^{-2}$ ) and floodplain lakes ( $141.29 \pm 175.23$  individuals  $100 \text{ m}^{-2}$ ) (Mann–Whitney *U*-test,  $n = 176$ ;  $P < 0.001$ ).

The river locations were characterized by shallow, relatively clear waters (compared to the floodplain lakes), a heterogeneous bed morphology with a variety of substrata (e.g. silt, mud, sand, gravel and pebbles), and the absence of aquatic and riparian vegetation. Here, a total of 26 fish species were caught (Table I), dominated by eurytopic and rheophilic species. *Pseudorasbora parva* were rarely caught ( $n = 16$  individuals, caught in nine of the 35 locations, from 152 sample sessions) and all catches consisted of adult specimens ( $\geq 24$  mm fork length,  $L_F$ ). The floodplain lakes were characterized by shallow ( $< 1$  m), relatively turbid waters, with little aquatic vegetation ( $< 0.1\%$  cover), but with an abundant shoreline vegetation (e.g. *Phragmites australis*, *Mentha aquatica*, *Myosotis scorpioides* and *Scirpus* sp.). Here, a total of 15 fish species were caught (Table I), dominated by eurytopic and limnophilic species. *Pseudorasbora parva* were frequently caught ( $n = 3058$  individuals, caught in seven of the 10 floodplain lakes, from 24 sample sessions), and catches consisted of both adults ( $\geq 24$  mm  $L_F$ ) and 0+ year larvae and juveniles ( $< 24$  mm  $L_F$ ). Since the floodplain lakes had not

TABLE I. Mean per cent composition of the fish fauna observed in the 35 river locations ( $n = 7690$  fishes) and the 10 floodplain lakes ( $n = 8774$  fishes), grouped according to flow preference (after Schiemer & Waidbacher, 1992)

Species	River locations	Floodplain lakes
Rheophilic A		
Brook lamprey <i>Lampetra planeri</i>	0.01	0
Nase <i>Chondrostoma nasus</i>	0.04	0
Dace <i>Leuciscus leuciscus</i>	0.04	0
Brown trout <i>Salmo trutta fario</i>	0.14	0
Barbel <i>Barbus barbus</i>	0.29	0
Minnnow <i>Phoxinus phoxinus</i>	2.03	0
Bullhead <i>Cottus gobio</i>	6.11	0
Chub <i>Leuciscus cephalus</i>	9.36	0
Rheophilic B		
Ide <i>Leuciscus idus</i>	0.05	0
Spined loach <i>Cobitis taenia</i>	1.38	0.10
Gudgeon <i>Gobio gobio</i>	8.69	0
Stone loach <i>Barbatula barbatula</i>	13.97	0.05
Eurytopic		
Eel <i>Anguilla anguilla</i>	0.22	0
Pikeperch <i>Sander lucioperca</i>	0.01	0
Ruffe <i>Gymnocephalus cernuus</i>	0.05	0
Silver bream <i>Blicca bjoerkna</i>	0.25	0
Pike <i>Esox lucius</i>	0.20	0.05
Bleak <i>Alburnus alburnus</i>	0.25	0.13
Bream <i>Abramis brama</i>	1.81	0.07
Roach <i>Rutilus rutilus</i>	21.42	0.07
Perch <i>Perca fluviatilis</i>	6.55	1.38
Carp <i>Cyprinus carpio</i>	0.05	10.86
Three-spined stickleback <i>Gasterosteus aculeatus</i>	25.97	49.75
Limnophilic		
Bitterling <i>Rhodeus sericeus</i>	0	0.06
Rudd <i>Scardinius erythrophthalmus</i>	0.07	0.11
Nine-spined stickleback <i>Pungitius pungitius</i>	0.83	0.84
Tench <i>Tinca tinca</i>	0	1.61
Topmouth gudgeon <i>Pseudorasbora parva</i>	0.21	34.85
Sunbleak <i>Leucaspis delineatus</i>	0	0.07

been flooded for at least 1 year, the presence of 0+ year fish indicated reproduction of *P. parva* in the floodplain lakes (Pollux & Korosi, 2002).

The results show that *P. parva* reached its highest densities in floodplain lakes, which acted as suitable spawning, nursery and adult habitats for this species. The low observed densities of predominantly adult individuals in the river locations furthermore suggest that the river merely serves as a dispersal corridor for adult specimens. Therefore, native species that use the floodplain lakes, at least during one of their life stages, are the most likely species to be affected by the presence of *P. parva*. These include limnophilic species, that reside

mainly in the floodplain lakes [e.g. bitterling *Rhodeus sericeus* (Pallas), rudd *Scardinius erythrophthalmus* (L.), tench *Tinca tinca* (L.), sunbleak *Leucaspius delineatus* (Heckel) and nine-spined stickleback *Pungitius pungitius* (L.)], as well as eurytopic species that use these floodplain lakes temporarily, for example, as a spawning and nursery habitat, and move to and from the floodplain lakes during periodic floods [e.g. roach *Rutilus rutilus* (L.), bream *Abramis brama* (L.), carp *Cyprinus carpio* L., perch *Perca fluviatilis* L., pike *Esox lucius* L. and three-spined stickleback *Gasterosteus aculeatus* L.]. Thus, although the impacts seem greatest in the floodplain lakes that are peripheral to the main channel, *P. parva* may also affect species in the main river channel.

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